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About TERI

A dynamic and flexible organization with a global vision and a local focus, TERI was established in 1974. While in the initial period the focus was mainly on documentation and information dissemination activities, research activities in the fields of energy, environment, and sustainable development were initiated towards the end of 1982. The genesis of these activities lay in TERI's firm belief that efficient utilization of energy, sustainable use of natural resources, large-scale adoption of renewable energy technologies, and reduction of all forms of waste would move the process of development towards the goal of sustainability.

The Bioresources and Biotechnology Division

Focusing on ecological, environmental, and food security issues, the Division's activities include working with a wide variety of living organisms, sophisticated genetic engineering techniques, and, at the grassroots level, with village communities. The Division functions through five areas—Centre for Mycorrhizal Research, Microbial Biotechnology, Plant Molecular Biology, Plant Tissue Culture, and Forestry/Biodiversity. The Division is actively engaged in mycorrhizal research. The Mycorrhiza Network has specifically been created to help scientists across the globe in carrying out research on mycorrhiza.

The Mycorrhiza Network and the Centre for Mycorrhizal Culture Collection

Established in April 1988 at TERI, New Delhi, the Mycorrhiza Network first set up the MIC (Mycorrhiza Information Centre) in the same year, and the CMCC (Centre for Mycorrhizal Culture Collection) – a national germplasm bank of mycorrhizal fungi – in 1993. The general objectives of the Mycorrhiza Network are to strengthen research, encourage participation, promote information exchange, and publish the quarterly newsletter, *Mycorrhiza News*.

The MIC has been primarily responsible for establishing an information network, which facilitates information sharing among the network members and makes the growing literature on mycorrhiza available to researchers. Comprehensive databases on Asian mycorrhizologists and mycorrhizal literature (RIZA) allow information retrieval and supply documents on request.

The main objectives of the CMCC are to procure strains of both ecto and VA mycorrhizal fungi from India and abroad; multiply and maintain these fungi in pure culture; screen, isolate, identify, multiply, and maintain native mycorrhizal fungi; develop a database on cultures maintained, and provide starter cultures on request. Cultures are available on an exchange basis or on specific requests at nominal costs for spore extraction/handling.



	New approaches	19
. 2	A simple staining technique for VAM fungi	18 18
	List of cultures available with the CMCC	19
11	Recent references	21
15	Forthcoming events	23
	11	 Estimation of total and living fungal biomass A simple staining technique for VAM fungi List of cultures available with the CMCC Recent references Forthcoming events

Role of mycorrhiza in tree plantings in the field. Part I. Field performance of mycorrhized plants

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Management of mycorrhizae and associated beneficial organisms in reforestation programmes comprises three major components-protection of indigenous soil communities, and evaluation of inoculation needs; integration of inoculation programmes into existing reforestation technology; and research. Adequate mycorrhiza formation is critical for ectomycorhizal trees growing on poor sites or in environments where seedlings must establish quickly to survive. Poor mycorrhiza formation reduces the chances of survival of seedlings and consequently, tree stocking which in turn leads to further loss of mycorrhizal inocula. Research is, therefore, needed on adaptation of mycorrhizal fungi to particular environmental factors; on interactions between tree seedlings and processes; and on the role of mycorrhiza and associated organisms in ecosystem structure and processes, particularly nutrient cycling, plant-plant interaction, and soil structure (Perry, Molina, and Amaranthus 1987).

Performance of mycorrhized plants on normal sites

Effect of mycorrhiza on survival rates

Studies conducted at the Forestry Research Institute, Seoul, Republic of Korea, on *Pinus rigida* seedlings inoculated with US strain of *Pisolithus tinctorius* on fumigated nursery soil and then transplanted in the field, showed that survival percentage of inoculated seedlings was 90% compared with 83% in controls (Lee, Park, and Lee 1988).

Inoculation of *Pinus resinosa* seedlings with *Hebeloma arenosa* in nursery significantly increased seedling survival following outplanting in the field but did not increase seedling growth in studies conducted at the Department of Plant Pathology, University of Wisconsin, Madison, USA. *H. arenosa* did not colonize roots growing from the root plug into the surrounding soil (MacFall and Slack 1991).

In studies conducted at the Department of Plant Pathology, Institute de Recerca i Tecnologia Agroalimentaries, Centre de Cabrils, Barcelona, Spain, bare-root Douglas fir seedlings inoculated with Laccaria laccata S-238 or uninoculated (infected by Thelephora terrestris or a Suillus sp.) in two Central France nurseries, were outplanted in a randomized block design at Pontavedra and in a completely randomized design at Garona. Survival after one growing season in the field differed considerably at the two reforestation sites, being much higher at Garona than at Pontavedra for all treatments. Within each site, survival of seedlings was not significantly related to either nursery source or inoculation treatment. Inoculated seedlings were taller than uninoculated seedlings irrespective of nursery source or site location and analysis of the data revealed that seedling growth after one growing season in the field was related to initial height (Alvarez, Parlade, and Pera 1990).

Studies conducted at Bioplanta, C P Campinas, SP, Brazil, showed that the number of surviving seedlings raised from heat treated, pregerminated seeds of cleopatra Citrus and inoculated with VAM in Styrofoam containers was 49168 from a total of 88960 seedlings (55.27%). Growth difference between the inoculated and non inoculated plants was much greater in fumigated soil than in non-fumigated soil where the uninoculated plants were colonized by native VAM fungi in two months (Lin, Mattos, Taveira, et al. 1987).

Studies conducted at the USDA (United States Department of Agriculture), Forest Service, Athens, GA, USA, showed that after eight years in the field, loblolly pine (*Pinus taeda*) with Pt (*Pisolithus*

^{*} Compiled from TERI database - Riza

tinctorius) index of 88 and 68 had better survival than nursery run control seedlings (Pt index 0) (Marx and Cordell 1987).

Studies conducted at the University of Wyoming, Department of Botany, Laramie, USA, on ectomycorrhiza formation, survivability, and physiognomic characteristics of conifer seedlings encountered one and two years post fire showed that *Pinus contorta* germinated and was abundant throughout first growing season and survival was about 50% for east-facing burn and edge and westfacing burn treatments. Abies lasiocarpa germinated during May and June but was rarely encountered by September. Few or no mycorrhizae were formed until September but by September all surviving seedlings were ectomycorrhizal. The number of ectomycorrhizae was positively correlated with the number of primary needles and the root-shoot ratio (Miller, McClean, Stantom, et al. 1998).

Effect on growth parameters

Outplanting performance of mycorrhiza inoculated seedlings was reviewed at the USDA, Forest Service, Pacific Northwest Research Station, Corvallis, Oregon, USA. The reviewed work included 568 unique combinations for 66 host plants (16 VAM, 49 ectomycorrhizal, and 1 VAM + ectomycorrhizal), 74 fungal species (8 VAM, and 66 ectomycorrhizal), and 5 inoculum types. Of these, 267 entries list improvement in one or more plant growth parameters, 41 list deleterious effect, 7 entries list both positive and negative effect, and 253 list no impact on seedling performance. In ectomycorrhizal entries, 41 host plants belonged to Pinaceae, 6 to Fagaceae and one each to Casuarinaceae and Leguminoceae. Vesiculararbuscular mycorrhizal host plants belonged to 16 families (Castellana 1990).

Studies conducted at the Department of Forest Resources, University of Minnesota, St Paul, USA, showed that the use of container culture and ectomycorrhizal inoculation with *Pisolithus tinctorius* significantly improved seedling root and shoot growth, water balance, and soil–plant liquid flow resistance of black oak seedlings outplanted under reforestation programmes. Seedlings planted without the benefit of ectomycorrhizal association may lack the growth potential necessary to compete with adjacent vegetation (Dixon, Garrett, Cox, et al. 1984).

Growth of *Pinus sylvestris* seedlings inoculated in the nursery with various forest mycorrhizal fungi (*Amanita muscaria, Lactarius rufus, Suillus variegatus, Tricholoma albobrunneum*, and an unknown fungus) was monitored for three years in field plantings in studies conducted at the Department of Mycology and Plant Pathology, Swedish University of Agricultural Sciences, Uppsala, Sweden. At outplanting, 10–40% of the root tips were mycorrhizal consisting of both the inoculated fungi and an indigenous nursery mycorrhizal fungus, *Thelephora terrestris*. Some of the inoculated seedlings were substantially smaller than the uninoculated seedlings. But after 2.5 years, seedlings in some of the treatments had up to 50% greater volume than the control seedlings. Even at low initial mycorrhizal colonization rates, some mycorrhizal species stimulated seedling growth. This was specially true for seedlings inoculated with *A. muscaria* — these were 50% smaller than control at the time of outplanting but grew rapidly to 20% larger than control 18 months after outplanting. This observation contradicts the general concept pertaining to the importance of seedling size at the time of outplanting (Stenstrom and Ek 1990).

Another study at the above university on *Pinus* sylvestris seedlings inoculated in the nursery with Laccaria laccata (isolate S 238 A, now known as L. bicolor), Hebeloma crustuliniforme (two isolates), and *Cenococcum geophilum*, showed that at the time of outplanting in the field 25–90% of the root tips were mycorrhizal with target fungi whereas 25–50% of the uninoculated control seedlings were spontaneously colonized by other fungi. Most of the inoculation in the nursery resulted in reduced seedling growth; growth rate inhibition was pronounced up to 1.5 years in the field except for seedlings inoculated with C. geophilum. By this time, seedlings inoculated with L. bicolor and H. *crustuliniforme* had 40–50% less volume than the control seedlings and the relative difference was maintained or was slightly decreased during following four years. The uninoculated nursery-produced seedlings had about 50% more volume than the corresponding control seedlings at outplanting. In the field, however, they grew relatively more slowly and consequently, were similar in growth to the control seedlings (Stenstrom 1990b).

Pinus spp., Eucalyptus spp., dipterocarps, and Casuarina equisetifolia were inoculated in the nursery with mycorrhizal tablets made from compressed mixture of basidiospores of *Pisolithus tinctorius* and Scleroderma cepa and then outplanted in the field in studies conducted at the Department of Forest Biological Sciences, University of Philippines. Inoculated seedlings showed an increase in height by 40-60%, diameter by 40-90%, and volume by more than 200% over the uninoculated plants. The positive response due to tablet inoculation in the nursery was observed even after three years in the field. In addition, mycorrhizal tablet inoculation was able to replace 60–85% of inorganic fertilizers required for the growth of pines and eucalypts in the field (Reynaldo and Cruz 1990).

Studies conducted at the Forestry Research Institute, Seoul, Republic of Korea, on *Pinus rigida* seedlings inoculated with US strain of *Pisolithus tinctorius* grown in fumigated nursery soil and then transplanted on exposed mineral soil, showed that height and root collar diameters of inoculated seedlings were 53% and 82%, respectively, more than those of uninoculated seedlings grown on nonfumigated soil. On outplanting, seedling growth (dry weight) was 20% higher in inoculated seedlings. Height increment after transplanting was 31% more in inoculated seedlings than in controls. The Korean strain of *P. tinctorius* was less effective than the US strain (Lee, Park, and Lee 1988).

Studies conducted at the USDA, Forest Service, Athens, GA, USA, showed that after eight vears in the field, loblolly pine with Pt (*Pisolithus tinctorius*) index of 88 and 68 had better growth than nursery-run control seedlings (Pt index 0). Volume and weight yield per acre were 50% greater for index 88 trees than controls. Distribution into four size classes showed that 67%, 81%, 80%, and 62% of index 88 trees were in the two largest classes of height, dbh, tree volume, and weight, respectively, compared to 53%, 63%, 59% and 39% of controls. Index 27, 46, and 68 trees were intermediate. Analysis also showed that average volume per acre positively correlated with Pt index more than 58. P. tinctorius can persist on a goodquality site and alleviate growth loss due to water deficit (Marx and Cordell 1987).

Comparative performance of mycorrhizal fungi

Studies conducted at the Heffley Reforestation Centre Ltd, Kamloops, British Columbia, Canada, showed that the fungi that established in nurseries are common in field plantings. In nurseries, the Engleman spruce is colonized only by *Thelephora* terrestris and Laccaria spp. under high nitrogen and phosphorus fertility but in low fertility, Amphinema byssoides and E. strain dominated. Mycelium radices atrovirens is also common at low fertility. On lodge pole pine, at low fertility levels, E. strain, Suillus sp., and Mycelium radices atrovirens are common. In field plantings, on normal reforestation sites, the fungi which produced significant increase in absolute or relative (annual increment as percentage of starting measurement) diameter growth as compared to control are Amphinema byssoides or E. strain on Engleman spruce, Suillus sp. on lodgepole pine, and *Rhizopogon* spp. on Douglas fir. The fungi which have been identified on spruce in Alberta and British Columbia of Canada include E. strain, Thelephora terrestris, Cenococcum geophilum, Amphinema byssoides, Mycelium radices atrovirens, Tommentella spp., Laccaria spp., Hebeloma like species, Tuber like species, and Suillus and Rhizopogon like species. The most abundant fungi on young seedlings are generally not members of Agaricales (Hunt 1990).

In studies conducted on *Pinus sylvestris* at the Department of Forest Mycology and Pathology, Swedish University of Agricultural Sciences, Uppsala, Sweden, non-aggressive forest fungi, especially *Amanita muscaria* and *Lactarius rufus*, tended to have the greatest benefits on seedling growth after outplanting in the field. Two-and-halfa years after outplanting, some of the inoculations resulted in seedlings with up to 40% more shoot volume that uninoculated controls. In contrast, rapidly colonizing, aggressive, and tenacious mycorrhizal fungi, especially *Laccaria laccata* and *Hebeloma crustuliniforme*, appeared to require too much of host's carbohydrate supply. Inoculations with these species, therefore, resulted in prolonged stunting. However, these fungi appeared to be better adapted to moist soils where they easily colonized pine roots. Inoculations with *L. laccata* and *H. crustuliniforme* also increased tolerance to drought stress (Stenstrom 1990a).

In studies conducted at the Institut Nationale Le la Recherche Agronomique, Equipe Microbiol Forestiere (INRA), Champenoux, France, nurserygrown Quercus petracea and Q. rubra seedlings were inoculated with Paxillus involutus or Hebeloma crustuliniforme or Laccaria laccata and at the end of first growing season, outplanted at two sites. P. involutus was the most competitive and efficient of the three fungi tested as its mycorrhizae were present seven years after outplanting and this markedly improved the growth of both oak species. In Q. rubra, growth stimulation was more marked during years with dry summer (Garbaye and Churin 1997).

Effect of fertilizers on mycorrhizal efficiency

In studies conducted at the Department of Plant Pathology, North Carolina State University, USA, loblolly pine seedlings were inoculated with 750 cm³ mycelium of each of *Pisolithus tinctorius*, Thelephora terrestris, and P. tinctorius plus T. terrestris or left uninoculated and fertilized with either nitrogen (ammonium nitrate at 145 kg/ha) or NPK (ammonium nitrate, calcium dihydrogen phosphate, and potassium chloride at 145, 50, and 100 Kg/ha, respectively). The seedlings were then planted on two sites an average site and a poor site. After four growing seasons on the average site, there was no difference in height, stem diameter, or survival among mycorrhizal fungal or fertilizer treatments. On the poor site, trees with T. terrestris and unfertilized were taller than those in other treatments. Trees with ectomycorrhiza formed by *P. tinctorius* and fertilized with nitrogen were 21% taller than those with T. terrestris or T. terrestris plus P. tinctorius ectomycorrhizae in the first year (Grand, Raleigh, and Krugner 1979).

Loblolly pine (*Pinus taeda*) seedlings were outplanted on eight reforestation sites following inoculation in the nursery with either *Pisolithus tinctorius* or *Thelephora terrestris* or naturally inoculated in the field and with two levels of ammonium sulphate fertilizer in studies conducted at the Southern Forestry Research Centre, Hot Springs, USA. The results showed that the seedlings with a high incidence of *P. tinctorius* mycorrhiza or fertilized with high rate of ammonium sulphate survived best. Trees with low levels of *P. tinctorius* mycorrhiza survived poorest. Growth after one year in the field was also related to nursery practice but subsequent growth appeared to be related to site quality and initial height. There appeared to be neither a treatment site interaction nor a carry-over effect beyond the first year (Mexal, Marx, and Morris 1979).

Effect of root pruning or suppression

Studies conducted at the USDA, Forest Service, Athens, GA, USA, showed that container-grown seedlings of loblolly pine and longleaf pine inoculated with *Pisolithus tinctorius* and planted on routine forest sites often failed to show improved field performance over uninoculated seedlings. However, when seedlings of both pines were chemically pruned with cupric carbonate in the greenhouse to alter root morphology and improve spread of *P. tinctorius* in egressed roots, they survived and grew better than uninoculated and untreated seedlings after two growing seasons (Ruehle 1987).

Shorea parviflora seedlings were collected nine months after logging, from beneath open, closed, and partially closed, canopies resulting from different logging treatments in studies conducted at the Institute of Terrestrial Ecology, Bush Estate, Midlothian, Scotland. Logging was done manually with minimal soil disturbance, compaction, or damage to the vegetation. Seedlings growing under closed canopy were smaller and had lower levels of mycorrhizal infection than those under open or partially closed canopies. Twenty-six ectomycorrhizal types were distinguished, and the greatest diversity was in seedlings under open or partially closed canopies. Based on their frequency (number of seedlings having a mycorrhizal type) and abundance (per cent infection of a mycorrhizal type if present), mycorrhizal types could be placed in two categories — those which occurred predominantly under the closed canopy and those which occurred both under open and partially closed canopies (Ingleby, Munro, Noor, et al. 1998).

Performance of mycorrhizal plants on difficult site

On degraded sites

Studies conducted at the Institut Nationale de la Recherche Agronomique, Laboratoire de Microbiologie Forestiere, Nancy, France, on Douglas fir (*Pseudotsuga menziesii*) seedlings inoculated naturally (with *Thelephora terrestris*) or artificially (with *Hebeloma cylindrosporum*) and planted on a Colluna healthland showed mortality rates of 13% and 31% respectively, three years after planting. Douglas fir seedlings inoculated naturally with T. terrestris or Suillus sp. or artificially with Laccaria *laccata* and planted on site following clear felling of a mixed coniferous broad-leaved stand, did not exhibit significantly different mortality rates but those inoculated with L. laccata had significantly greater height growth after one year (about 115 vs 60 cm) (Tacon, Bouchard 1991).

Vesicular-arbuscular mycorrhizal fungi including *Glomus fasciculatum*, *G. intraradices*, *G.* dimorphicum, Scutellospora calospora, and S. gigantea improved the growth of hardwood tree seedlings particularly on degraded alkaline soils according to studies conducted at the Biomass Research Centre, National Botanical Research Institute, Lucknow, India (Behl 1990).

In studies conducted at the Great Lakes Forestry Centre, Forestry Canada, Ontario, Canada, 14-week old nursery-grown seedlings of black spruce (Picea mariana) and jackpine (Pinus banksiana) inoculated with fragmented hyphae of one of the five ectomycorrhizal fungi — Hebeloma cylindrosporum, Pisolithus tinctorius, Laccaria bicolor, L. proxima, or Rhizopogon rubescens — were outplanted on two sites each, namely black spruce on peatland site and stony loam site and jack pine on stony loam site and sandy site. Inoculation of black spruce with L. proxima resulted in significantly better shoot growth on both sites and in higher foliar concentration of potassium and zinc on peatland site, and with H. cylindrosporum resulted in improved height growth on the peatland site after two years as compared to controls. Black spruce seedlings inoculated with L. bicolor were significantly smaller than uninoculated seedlings. This difference in size persisted for two growing seasons. Inoculation of jack pine with L. proxima resulted in significantly better shoot growth on both sites and significantly higher foliar concentrations of nitrogen, phosphorus, potassium, and zinc at stony loam site after two years. Inoculation with H. cylindrosporum and L. bicolor improved height growth of jack pine at stony loam site after two years over control plants. P. tinctorius did not form mycorrhiza on both species (Browning and Whitney 1992).

On Disturbed sites

Studies conducted at the Institute of Mycological Research and Development, USDA, Forest Service, Southeastern Forest Experiment Station, Forestry Science Laboratory, Athens, showed that forestation of routine planting sites and revegetation of drastically disturbed lands such as acid coal spoils, Kaolin spoils, burrow pits, and severely eroded sites in the eastern USA, have been significantly improved by planting tree seedlings tailored in the nursery by the ectomycorrhizal fungus Pisolithus *tinctorius*. *P. tinctorius* mycorrhizae improve seedling survival and growth on good-quality reforestation site especially during drought. On disturbed sites, P. tinctorius promotes the assimilation and absorption of nutrients, increases water absorption, reduces the absorption of toxic elements, and apparently reduces the overall stress inherent in these sites (Marx and Cordell 1998). Soils were prepared by total ploughing to 50-cm depth following removal of stumps or top humus layers contaminated with heavy metals or by scarification of the site following preparation of individual planting holes with an auger in studies conducted at the Katedra Fitopatologii Lesnej

Akademia, Rolnicza, Krakow, Poland. The site was prepared by the removal of 90-year-old *Pinus* sylvestris stand from a moist coniferous site with average annual deposits of 0.61 mg/m³ sulphur dioxide and 110.4 tonne/m² dust. Planting was carried out with two-year-old plants of *Pinus* sylvestris, *P. nigra*, *P. strobus*, *Larix decidua*, *Alnus* incana, Betula verucosa, and Quercus rubra. Mycotrophy (living mycorrhizae as percentage of total number on 30 lengths of roots, each length of 3 cm) after six years was 13%–21% greater in total ploughed site than in individual planting hole site (Kowalski 1987).

Studies conducted at the USDA, Pacific Northwest Experiment Station, Corvallis, USA, highlighted investigations on the ability of soils to maintain viable population of ectomycorrhizal fungi following forest disturbances. Ectomycorrhizal formation on seedlings generally decreases as time between disturbance and reforestation increases because survival of propagules and input of spores of ectomycorrhizal fungi appear insufficient to maintain mycorrhizal potential of disturbed sites in the absence of host plants. Also, degree of organic matter lost from a site can influence populations of mycorrhizal fungi. Mycorrhizal formation was reduced by 90% and basal area by 43 % on Douglas fir seedlings grown in soils where forestry practices drastically reduced organic matter levels. Rapid mycorrhiza formation and improved seedling performance on Douglas fir seedlings in association with white leaf manzanite (Arctostaphylos viscida) were found comparable to adjacent annual grass areas. In another study, number and rapidity of ectomycorrhizal formation of Douglas fir seedlings increased with proximity to clumps of green-leaf manzanite. Also, ectomycorrhizal fungi supported by non-coniferous hosts can actively colonize feeder roots of coniferous seedlings (Amaranthus and Perry 1990).

Studies conducted at the University of Calgary, Department of Biological Sciences, Calgary, Canada, on ectomycorrhizal and arbuscular mycorrhizal response to clear-cutting and the application of 5-and 10-cm deep layers of chipped aspen (Populus tremuloides) wood in a boreal, mixed wood site dominated by aspen poplar showed that there was extensive mortality of aspen roots during first two years after cutting although many roots persisted for one year and small pockets of regenerating roots were detected after two years. Mycorrhizal colonization of aspen roots was 90-100% in all treatments indicating that inoculum potential of the ectomycorrhizal fungi was maintained for two years post-harvest. Cenococcum geophilum, Cortinarius spp., Russula spp., and Tomentella spp. were abundant on roots in all treatments and clear-cutting and wood chip application did not reduce the regeneration potential of these fungi. *Piloderma byssinum* occurred frequently on unharvested plots while Hebeloma was abundant in plots with 10-cm wood chips and Mycelium radices

atrovirens in plots with 5-cm wood chips (Visser, Maynard, and Danielson 1998).

In studies conducted at the USDA, US Forest Service, S Main, Moscow, USA, western white pine and Douglas fir were outplanted on two sites (low altitude, relatively harsh site and a higher altitude, more moderate site) and were given four treatments — mounding surface horizons with competing vegetation, mounding with physical or chemical control of competing vegetation, scalping to control competing vegetation, and control (with no treatment). Mounding with no competing vegetation produced small seedlings (5–15 g for both species) with low numbers of ectomycorrhizal roots. Mounding with competing vegetation produced large seedlings (20-48 g) with moderate number of short roots. Scalping produced small seedlings (8–16 g) with high number of short roots. Controls produced small seedlings (8–13 g) with moderate number of short roots. Douglas fir produced more short roots on harsh site than on moderate site, high root-shoot ratio (above 1.0) on mounds with no competing vegetation on harsh site and on mounds with competing vegetation moderate site. Western white pine produced more ectomycorrhizae more rapidly than Douglas fir, produced high root-shoot ratio (above 0.60) in mounds with competing vegetation on harsh site (Harvey, PageDumroese, Jurgensen, et al. 1996).

On mined sites

Studies conducted at the Native Plants Inc., Salt Lake City, UT, USA, on field trials of Salix viminalis and S. dasyclados on a mined peatland site revealed statistically significant growth stimulation due to mycorrhiza formation after mycorrhizal inoculation after one growing season. In greenhouse, both species formed ectomycorrhizae with Amanita spp., Cortinarius purpurascence, Entoloma nidorosum, Entoloma spp., Hebeloma crustuliniforme, H. pusillum, Laccaria bicolor, and Paxillus involutus. In field trials, plants inoculated with *Entoloma* spp. were sometimes twice as large as control plants. However, such effects were observed only in plants receiving normal phosphate fertilization as opposed to low phosphate application and were not consistent from one season to another. With the inoculum of other species (Cortinarius, Hebeloma, and *Paxillus*), some evidence of growth enhancement was found in the field but these results were sometimes attributable to non-symbiotic effects of inoculation (Loree, Lumme, Neimi, et al. 1989).

One-year-old bare-rooted seedlings of loblolly pine (*Pinus taeda*), artificially inoculated with *Pisolithus tinctorius* and uninoculated controls (naturally infected with *Thelephora terrestris* and ectomycorrhiza), were outplanted on a surface mine site which had previously been contoured and hydroseeded with a mixture of herbaceous ground cover species in studies conducted at the Department of Range, Wild Life and Forestry, University of Nevada, Reno, USA. On half of each of mycorrhizal plants, NPK fertilizer (336 kg/ha) was broadcast. After seven years, survival and growth of *P. tinctorius* inoculated trees were significantly improved relative to control. Fertilizer treatment elicited a significant reduction in survival and gave marginal growth responses in trees of both mycorrhizal treatments. Also, *P. tinctorius* infected seedlings had more nitrate and less zinc in their needles than control (Walker, West, McLaughlin, et al. 1989).

In studies conducted at the Division of Reclamation, Ohio Department of Natural Resources, Columbus, Ohio, USA, the success of reforestation programme of abandoned mined land on sites of 2-75 acres is attributed to the use of tree seedlings inoculated with ectomycorrhizal fungus, Pisolithus tinctorius. Traditional practices for reclaiming abandoned mined lands are expensive (\$7000 per acre). Reforestation is, therefore, suggested as a costeffective and low-maintenance alternative (\$300 per acre) for reclaiming strip mines not eligible for traditional reclamation. Inoculated Virginia pine (Pinus virginiana) planted on barren shales (pH 2.8) had 99% survival, with mean diameter at breast height of 2.9 cm and mean height of 120.9 cm in February 1989 (Caldwell 1990).

Loblolly pine seedlings were transplanted on a pyretic coal mine site in a commercial reclamation in effort studies conducted at the Department of Plant Pathology and Horticulture and Landscape Horticulture, University of Kentucky Lexington, Kentucky, USA. After five years, the trees which became naturally infected with *Pisolithus tinctorius* or other ectomycorrhizal fungi which produce sporocarps had twice the height and stem diameter of the trees not associated with sporocarps. Soil pH was lower in the root zone of trees associated with P. tinctorius than in those not associated with P. *tinctorius* and sulphate was higher. No differences were found in ferric or ferrous ions or in population of two bacterial species, Thiobacillus ferrooxidance and T. thiooxidans, which catalyze production of sulphuric acid from pyrite. Apparently, P. tinctorius benefits its hosts by mechanism other than inhibition of Thiobacillus spp. (Hendrix, Hunt, and Maronek 1985).

In studies conducted at the Department of Environmental Biology, University of Guelph, Guelph, Ontario, Canada, seedlings of yellow poplar (Liriodendron tulipifera) were raised in nursery beds with all combinations of four mycorrhizal treatments — no inoculum, Glomus spp., G. mosseae, and natural stand soil — and four fertilizer treatments — no fertilizer, base (15:15:15 NPK, 560 Kg/ha), base plus two urea top dressings (39:0:0 NPK, 112 kg per ha), and base plus four urea top dressings - for one year. Nursery-raised plants were then outplanted at three locations namely a reclaimed stripmined site, an eroded old field site (shallow to shale bed rock), and river terrace site. Fertilizer treatment significantly affected the size of the seedlings in the nursery but differences were smaller in the outplantings after two years. Mycorrhizal inoculation did not greatly affect seedling size in the nursery but affected postplanting height growth. Seedlings inoculated with *Glomus* spp. grew the most, followed by seedlings inoculated with natural soil, G. mosseae, and no inoculum treatments. Seedling survival and foliar concentration of nitrogen and phosphorus were not significantly impacted by treatments in the nursery. Seedling growth was best at the river terrace site and worst at the eroded field site. Difference in seedling growth on river terrace site due to site quality overshadowed differences due to mycorrhizae and fertilizer treatments. Both Glomus inoculated and natural soil inoculated seedlings grew better under harsher conditions (Williams and David 1993).

Seedlings of Pinus contorta, P. flexilis, and Pica engelmannii, colonized with Pisolithus tinctorius, Suillus granulatus, and Cenococcum graniforme or uninoculated (colonized by a wild fungus) in nursery, were outplanted on a high-altitude (3200 m) molybdenum tailing waste site covered with deep mine waste material in studies conducted at the Department of Forest and Wood Science, Colorado State University. Fort Collins, USA. The seedlings were given four fertilizer treatments – no fertilizer, 90 kg/ha phosphorus, 90 kg/ha phosphorus + 68 kg/ha nitrogen, and sewage sludge and wood chips at 4500 kg (dry weight) per ha. Combined mortality after two growing seasons, was 26.1%. There was significantly greater mortality in sewage sludge and wood chip treatment in both pines. Effect of mycorrhizal treatment on mortality was not significant. Height growth in all the three tree species was greatest with S. granulatus and the wild fungus while the height growth of seedlings infected with *P. tinctorius* was generally about half the growth of the best fungal treatment (Steven and Reid 1979).

On basalt soils

Studies conducted at Northern Arizona University School of Forestry, Flagstaff, Arizona, USA, on *Pinus ponderosa* grown in containers, inoculated in the nursery with mycelium of *Suillus granulates*, and outplanted along with uninoculated control on a basalt-derived soil during the middle of growing season showed that at the end of growing season, there had been no significant growth in either control or inoculated seedlings. However there was 70% survival in the controls as compared to 93.3% in mycorrhizal seedlings (Carnett Z J 1979).

Performance of mycorrhizal bare-rooted vs. container-grown seedlings

Research at the INRA, Centre de Recherches Forestieres de Nancy, Champenoux, Seichamps, France, shows that more than 80% of the forest trees are planted bare-rooted because the production cost of such plants is lower than that of the container plants. When lifted from the nursery, the plants loose in the soil a large proportion of their root system, especially the apical parts, the mycorrhizae. Furthermore, all the connections between the soil and the plant through external mycelium and mycelial strands are destroyed. Then storage of plants before planting modifies mycorrhizal metabolism. During studies at the Institute, modification of the polypeptide profile as well as decrease in respiration have been recorded and related to mycorrhizal population evolution on the root system during the first year after plantation. Different types of ectomycorrhizae respond differently to transplantation stress. Some ectomycorrhizal fungi can thus be selected according to this criterion for controlled inoculation in the nursery (Al-Abras, LeTacon, and Lapeyrie 1988).

Lateral root development

In studies conducted at the USDA, Forest Service, Institute for Mycorrhizal Research and Development, Forestry Science Laboratory, Athens, USA, bare-rooted and container-grown loblolly pine (Pinus taeda) seedlings with Pisolithus tinctorius mycorrhiza were grown for 22 weeks and excavated at four-week intervals to study the development of lateral root system. The result showed that the pattern of lateral root growth from original root system differed between bare-root and containergrown seedlings. Bare-root seedlings produced 60% more laterals and had significantly more horizontal lateral root egress from the middle horizontal zone of the original root system than the container grown seedlings. Bare-root seedlings also had significantly more Pisolithus tinctorius mycorrhizae on egressed roots and the mycorrhizae were found at significantly greater distances from the original root system than those of containergrown seedlings after 22 weeks (Ruehle 1985).

Effect on survival

Studies conducted at the D L Phipps State Forest Nursery, Elkton, USA, showed that bare-root seedlings of Douglas fir (*Pseudotsuga menziesii*) survived and grew better during the first year than container-grown stock on a draughty site. After three years, survival still differed significantly but height growth did not. Shading improved survival and growth. Application of liquid suspension of spores of *Pisolithus tinctorius* was ineffective and no mycorrhizae developed from this fungus (Pilz and Znerold 1989).

In studies conducted at the Department of Forest Science, Texas A and M University, College Station, Texas, USA, survival of untreated bareroot seedlings, bare-root seedlings treated with Terra-Sorb root dip (a hygroscopic starch), container seedlings, and container seedlings inoculated with *Pisolithus tinctorius* was compared for two sources of loblolly pine (*Pinus taeda*) including one drought-resistant source and one source of short leaf pine (*P. echinata*) at 2, 6, 9, 14, and 21 months after planting on three drought-prone sites in Northeast Texas. Bare-root seedlings survived better than all container seedlings on the most severe site. On a more moderate site, container seedlings (without mycorrhizae) showed maximum survival and followed treated bare-root seedlings and then untreated bare-root seedlings (Echols, Meier, Ezell, et al. 1990).

Effect on growth parameters

Studies conducted at the School of Forestry, Fisheries and Wildlife, University of Missourie, Columbia, USA, on black oak (*Quercus velutina*) seedlings outplanted on two clearcut sites showed that container-grown seedlings exhibited height and diameter growth superior to bare-root stock, and mycorrhizal plants were better than nonmycorrhizal plants during the first two to three seasons. Thereafter, container-grown and bare-root seedlings, and mycorrhizal and non-mycorrhizal seedlings grew at comparable rates such that the differences in height and diameter were greatly reduced six years after outplanting (Parker, Moorhead, Pallardy, et al. 1985)

Effect on photosynthesis and water potential

In further studies conducted at the above university, northern red oak (Quercus rubra) containerized one-year-old stock inoculated with Pisolithus tinctorius and two-year-old bare-root, uninoculated seedlings were planted in two contiguous high quality oak sites each of which was divided between a shelter wood plot and clear-cut plot. Half of the seedlings were clipped at approximately 8 cm above the root collar. During the first year after over storey cuttings, seedlings in the shelterwood plot had significantly lower water potentials, photosynthetic and transpiration rates, and stomatal conductance than seedlings in the clearcut plot. The greater level of seedling water stress under the shelter wood was associated with restricted root system development and possibly greater competition from over storey trees for water than in the clearcut plot where herbaceous vegetation was just becoming established. Containerized seedlings had significantly higher seasonal average midday water potentials than did the bare-root seedlings following outplanting but both had similar photosynthetic and transpiration rates and stomatal conductance. Clippings of seedlings did not have a consistent effect on seedling growth or physiological responses during the first year (Crunkilton, Pallardy, and Garrett 1988).

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Research findings

Arbuscular mycorrhizal dependency of Eucalyptus tereticornis Sm: how real is it?

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Introduction

Most hosts of non-arbuscular mycorrhizal fungi never establish functional arbuscular mycorrhizal symbiosis in nature (Giovannetti and Sbrana 1998). Among the known ectomycorrhizal hosts, eucalyptus is an exception as it forms arbuscular mycorrhiza also (Malajczuk, Lindermann, Kough, et al. 1981, Lapeyrie and Chilvers 1985). Which of the eucalyptus mycorrhizae is functional: ectomycorrhiza or arbuscular mycorrhiza, or both? Results of some of the arbuscular mycorrhizal dependency or inoculation response studies of Eucalyptus (Bhat Narayan, Jayarajan, and Ramraj 1993, Sharma, Chauhan, and Adholeya 1999) are not convincing to answer the question. Results of this study on arbuscular mycorrhizal dependence of *Eucalyptus tereticornis* in a low-phosphorus alfisol suggest that the species may not be economically dependent on arbuscular mycorrhiza (Trappe 1989) in phosphorus-deficient Indian soils although its roots may be substantially infected by arbuscular mycorrhizal fungi.

Materials and methods

Dependency study was carried out twice in a nursery house using protocol for estimation of inoculation response of *E. tereticornis* seedlings against graded available phosphorus in transported laterite soil (coarse loamy Typic Haplustalf) having pH 6.0, organic carbon 0.27%, total nitrogen 0.03%, and available phosphorus and potassium 5.0 and 13.5 ppm, respectively. Available phosphorus status of the soil was raised to 15, 25, 50, and 100 ppm with single superphosphate by the usual process of incubation-equilibration for two months.

Maize root based mixed arbuscular mycorrhizal species (*Glomus mosseae*, *G. fasciculatum*, and *G*.

aggregatum) effective inoculum developed in the usual manner was used to raise the most probable number of infective propagules of the treatment soil from 8 to 75 per gram. Mycorrhizal efficiency of the inoculum was tested against standard indicator plants. Heat-killed root inoculum in equivalent amount was used in the control set. Soil, plant, and mycorrhizal analyses were done following standards methods. Dependency was estimated as gain in efficiency of mycorrhizal plants in yield parameters at individual phosphorous levels (Powell and Daniel 1978, Plenchette, Fortin, and Furlan 1983) and the same estimate was used to compare mycorrhizal response of all parameters at any phosphorus level. Data presented are average of two experiments carried out during September-March 1996/97 and 1997/98.

Results and discussion

E. tereticornis seedlings raised in sterile sand were infected by arbuscular mycrorrhizal inoculation in steamed soil resulting in 90% root infection intensity at 100 days. Root-shoot dry weight ratio of the infected seedlings in inoculated soil increased significantly but this could not be concluded as 'atypical mycorrhizal response' as the inoculum used was crude arbuscular mycorrhiza infected root mass (Table 1).

Seedlings transplanted in normal treatment soil against graded phosphorus with added infected root inoculum showed unexpected response patterns. Irrespective of the levels of phosphorus at 110 and 240 days, mycorrhiza-inoculated plants responded negatively as compared to uninoculated plants, for both plant height and shoot dry matter yield, although phosphorus response of these parameters was positive, particularly at high levels

Table 1 Response of Eucalyptus tereticornis seedlings to arbuscular mycorrhizal inoculum in steamed soil at 100 days

Parameters	Steamed soil	Steamed soil + root inoculum	Estimated t ^a
Infective inoculm status, MPN per g	Nil	75	_
Plant height (cm/plant)	35.0	40.2	3.68
Root dry weight (g/plant)	0.127	0.297	3.74
Shoot dry weight (g/plant)	0.261	0.393	3.35
Root-shoot dry weight ratio	0.49	0.76	_
% root colonization	0	90.0	_
Shoot phosphorus content (mg/plant)	0.18	0.47	_

^a $t_{0.05}$ value at 18 d.f = 2.21

(Tables 2 and 3). Except at 5- and 15-ppm phosphorus levels, plant height and shoot dry weight of mycorrhiza-inoculated plants at 110 days were less than that of uninoculated plants where the negative effect for shoot weight was significant at high soil phosphorus level. The positive shoot weight yield response of the inoculated plants at 110 days 5and 15-ppm phosphorus levels, although translated to about 30% inoculation efficiency or relative mycorrhiza dependency, was only marginal and was terminated at 240 days (Table 3).

Root weight yield response to arbuscular mycorrhizal inoculum was different from that of shoot (Tables 2 and 3). Irrespective of the soil phosphorus level and at all levels, except 50 and 110 ppm, root dry weight of the inoculated plants at 110 days was significantly higher than that of the uninoculated plants. The difference was most pronounced at low phosphorus levels where shoot dry weight had shown a marginal positive response. At 240 days, this reversed as the root dry weight of inoculated plants. Characteristically, there was a corresponding decline in shoot weight also and the mycorrhizal inoculation efficiency or dependency estimate for shoot weight became negative even at the lowest phosphorus level (5 ppm). Both phosphorus concentration and content in shoot at 240 days, however, showed a predictable increase with inoculation where efficiency declined typically at high soil phosphorus levels (Figure 1). There was significant gain in root infection intensity due to inoculation, which also declined at high phosphorus levels.

Net change in whole plant dry matter due to inoculation was positive at 110 days till 15 ppm soil phosphorus but declined sharply with further increases, to get neutralized at about 60 ppm and negative at higher levels (Figure 2a). At all phosphorus levels, the incremental change in whole plant dry matter of inoculated plants at 110 days, partitioned more to roots than stems and above 20 ppm, the roots gained in dry matter apparently by depletion of shoots in addition to the gain of whole of the incremental dry matter due to inoculation (Figure 2a). At 240 days, the negative response in whole plant dry matter production due to inoculation was even more pronounced and advanced to very low soil phosphorus level (Figure 2b). The negative response at 240 days partitioned almost equally between root and shoot, particularly at high phosphorus levels.

	Plant height (cm)			Shoot d	Shoot dry weight (g)			Root dry weight (g)		
Phosphorus levels	UIª	l ^b	MIE℃	UI	I	MIE	UI	I	MIE	
P ₀ (5 ppm)	40.4 ^{cd}	40.8 ^{CD}	1.0	0.29 ^D	0.44 ^{BCD}	34.1	0.08 ^c	0.38 ^{AB}	78.9	
P ₁ (15 ppm)	41.3 ^{CD}	42.7 ^c	3.3	0.38 ^{BCD}	0.54 ^{ABC}	29.6	0.14 ^{BC}	0.52	73.1	
P ₂ (25 ppm)	55.0 ^A	43.2 ^D	-27.3	0.57 ^{ABC}	0.51 ^{ABC}	-11.8	0.16 ^{BC}	0.46	65.2	
P ₃ (50 ppm)	51.3 ^{AB}	41.4 ^{CD}	-23.9	0.61 ^{AB}	0.47 ^{ABCD}	-29.8	0.20 ^{ABC}	0.41	51.2	
P ₄ (100 ppm)	48.8 ^B	41.1 ^{CD}	-18.7	0.68	0.45 ^{BCD}	-51.1	0.29 ^{ABC}	0.32 ^{AB}	9.4	
Mean	47.4	41.8		0.51	0.48	-6.3	0.17	0.42	59.5	
S. Em ±		1.71			0.07			0.07		
LSD _{0.05} (M x P)		4.88			0.19			0.21		
S.Em ±		0.76			0.03			0.03		
LSD _{0.05} (M)		2.16			0.08			0.08		

Table 2Response of Eucalyptus tereticornis seedlings to arbuscular mycorrhizal inoculation in an alfisol against graded phosphorus at 110 days

^a Uninoculated, ^b Inoculated, ^c Mycorrhizal inoculation effect

 Table 3
 Response of Eucalyptus tereticornis seedlings to arbuscular mycorrhizal inoculation in an alfisol against graded phosphorus at 240 days

	Plant height (cm)			Shoot dry weight (g)			Root dry weight (g)		
Phosphorus levels	UIª	l ^b	MIE ^c	UI	Ι	MIE	UI	Ι	MIE
P _o (5 ppm)	63.8 ^H	65.7 ^{FG}	2.9	2.36 ^E	2.28 ^E	-3.5	2.64 ^D	2.80 ^{CD}	5.7
P ₁ (15 ppm)	65.6 ^{FG}	74.8 ^D	12.3	2.66 ^{DE}	2.65 ^{de}	-0.4	3.82 ^{BC}	2.70 ^D	-41.5
P ₂ (25 ppm)	77.0 ^c	72.4 ^E	-6.3	3.96 ^c	3.02 ^{DE}	-31.1	3.67 ^{BC}	2.60 ^D	-41.1
P_{3}^{2} (50 ppm)	81.2 ^B	68.4 ^{EF}	-18.7	4.35 [₿]	3.04 ^{DE}	-43.1	4.43 ^B	2.97 ^{CD}	-49.2
P ₄ (100 ppm)	86.14	66.6 ^F	-29.3	5.22	3.12 ^D	-67.3	5.53*	3.56 ^c	-55.3
Mean	76.7	69.6	-10.2	3.71	2.82	-31.6	4.02	2.93	-37.2
S. Em ±		0.45			0.17			0.29	
LSD _{0.05} (M x P)		1.28			0.49			0.84	
S.Em ±		0.20			0.08			0.13	
LSD _{0.05} (M)		0.57			0.23			0.37	

^a Uninoculated, ^b Inoculated, ^c Mycorrhizal inoculation effect

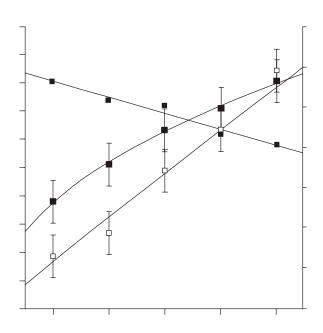


Figure 1 Shoot P content of mycorrhizal and non-mycorrhizal *E.tereticornis* plants at 240 days and gain in root infection intensity at 200 days due to inoculation

The observed negative mycorrihzal inoculation response of eucalyptus, even at a relatively low soil phosphorus level, may not be a unique case (Koide 1985) and can be explained by the high sink demand of mycorrhizal roots for photosynthate (Marschner 1995). Significantly, Sharma, Chauhan, and Adholeya (1999) reported almost similar results with a similar soil using phosphorus levels even lower than that used in this study. They reported 13%-26% dependency between 0.67 to 10 ppm Olsen's phosphorus and negative response at phosphorus levels higher than that at 112 days. The present study reports about 30% dependency at 110 days and negative dependency at 240 days between 5 and 15 ppm Olsen's soil phosphorus. It is doubtful whether such small dependency

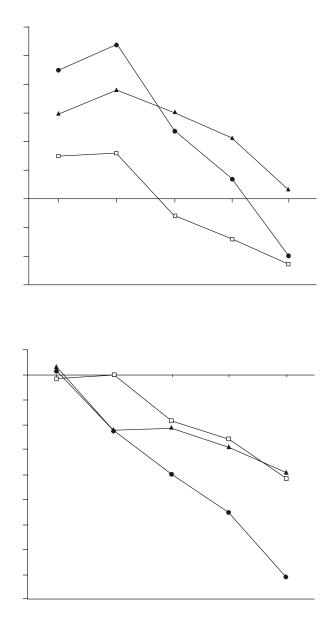


Figure 2 Partition of net change in dry matter content of AM inoculated *E. tereticornis* plants between shoot and root; (a) 110 days, (b) 240 days

estimates, as above, can be interpreted as significant (Sharma, Chauhan, and Adholeya 1999) particularly when the incremental shoot yield response even under management ambience of nursery culture is only marginal under strong phosphorus stress. Seedlings, 110–112 days old, may be too young to reflect arbuscular mycorrhizal dependency of a tree plant.

Non-mycorrhizal eucalyptus plants at 240 days responded to phosphorus by higher shoot weight formation apparently due to a parallel increase in root development. The response changed in mycorrhizal plants and there was a strong and significant decline in root weight resulting in a corresponding decline in shoot weight. This happened inspite of a significant gain in phosphorus acquisition in mycorrhizal plants where the gain apparently did not transform to increased shoot weight. Graw, Moawad, and Rehm (1979) reported a similar observation in sunflower (*Helianthus annuus*) which showed no response to inoculation in spite of strongly developed mycorrhiza and increased phosphorus uptake. Such negative mycorrhiza inoculation response in shoots would be the most logical response of a tree plant without any pronounced phosphorus deficit for growth but with limited photosynthetic capacity at early developmental stage to satisfy the extra cost of mycorrhization (Sanders 1993). The marginally positive mycorrhiza response at early seedling stage, at low phosphorus levels, may be due to the observed increase in root development (Figure 3) which has been often seen in inoculated tree seedlings casuarina and eucalyptus (Srinivas, Shanmugam, and Ramaraj 1989), Acacia (Sharma, Gaur, Bhatia, et al. 1995), Dalbergia (Agarwal and Chauhan, 1995), and E. tereticornis (Sharma, Gaur, Bhatia et al. 1995). Thus we conclude that *E. tereticornis* may not be an arbuscular mycorrhizal dependent tree plant at early developmental stage, although it is infected by arbuscular mycorrhizal fungi freely, improving its phosphorus acquisition efficiency in

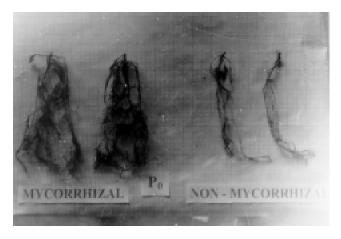


Figure 3 Root growth response of *E. tereticornis* to AM inoculation at 110 days as compared to no inoculation

low-phosphorus soil. What happens at later stages needs to be seen.

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Effect of different rice (*Oryza sativa* L.) based cropping systems on population dynamics of native vesicular-arbuscular mycorrhizal fungi under rainfed upland ecosystem

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Introduction

The association of rainfed upland rice, grown under aerobic soil conditions, with native VAM (vesicular-arbuscular mycorrhizal) fungi has been reported (Maiti, Variar, and Saha 1995). Augmenting native VAM fungi population under rainfed upland rice agro-ecosystem, inherently poor in soil nutrient status, could enhance the advantage of such association. The present investigation evaluated different rice-based cropping systems which could augment native VAM fungi population in soil. The agro-ecosytem under study was predominantly monocropped with wet-season (June-October) rice.

Materials and methods

A long-term fixed plot experiment was conducted during wet seasons of 1993–97 in the experimental farm of Central Rainfed Upland Rice Research Station, Hazaribag. The four cropping systems evaluated were

- 1 CS₁: rice sole (1993–97)
- 2 CS₂: rice followed by toria (*Brassica campestris* L. var. *toria*) during 1993 and 1994, rice followed by horse gram (*Dolichos biflorus* L.) during 1995,

maize (Zea mays) followed by niger (Guizetia obyssinica L. f. Cass) during 1996, and rice sole in 1997

- 3 CS₃: rice and red gram (*Cajanas cajan* L.) intercropping (1993–97)
- 4 CS₄: farmers' practice of crop rotation—rice sole in 1994, fallow in 1995, black gram (*Phaseolus mungo* L.) in 1996, and rice sole in 1997.

There were two sets of the experiment. The fallow period was subjected to three or four OSP (off-season [November to May next year] ploughing) in one set and NoOSP (without OSP) in the second. This amounted to eight treatments— T_1 : $CS_1 + OSP, T_2$: $CS_1 + NoOSP, T_3$: $CS_2 + OSP, T_4$: $CS_2 + NoOSP, T_5$: $CS_3 + OSP, T_6$: $CS_3 + NoOSP, T_7$: $CS_4 + OSP, T_8$: $CS_4 + NoOSP$.

Except rice, other crops were local varieties being used by the farmers. During 1993/94, a local variety (brown gora) of rice was used which was replaced by an improved variety (Vandana) during 1995–97. Fertilizers were applied as follows

- ⁿ CS_1 and $CS_3 = 40:30:20$ (N:P₂O₅:K₂O in kg/ha)
- ⁿ $CS_2 = rice: same as CS_1; maize: 90:60:40$ (N:P₂O₅:K₂O in kg/ha)

ⁿ CS_4 = rice: same as CS_1 ; BG: no fertilizer.

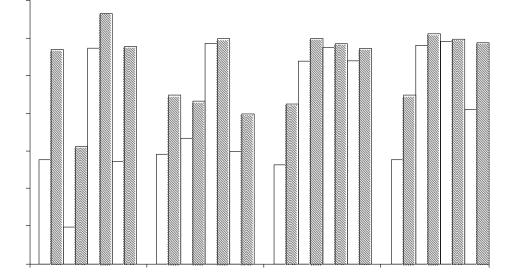
The experiment was conducted in fixed, large $(10'12 \text{ m}^2)$ unreplicated plots. Infective propagules of native VAM fungi in soil were quantified using the MPN (most probable number) method (Porter 1976). Three samples (each consisting of five pooled sub-samples) from each plot were analysed at two stages—at the beginning of the crop season (June) and at the end of the crop season (September end) during 1995–98.

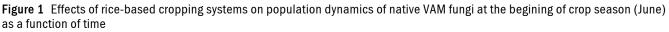
Results and discussions

In 1993, when the experiment was initiated, all plots had almost uniform native VAM fungi population (13 infective propagules per gram of soil) (Maiti, Variar, Singh, et al. 1997). By June 1995, different treatments resulted in varied populations ranging from 24 per gram of soil for T_1 to 210 per gram of soil for T_6 (Figure 1).

Highest residual positive effect on multiplication of VAM fungi population was observed in T_6 in June 1995 (Figure 1) but T_6 did not produce any substantial effect over T_5 in subsequent cropping seasons (June 1996–98). On the other hand, T_2 , T_4 , and T_8 increased native VAM fungi population over their OSP counterparts (T_1 , T_3 , and T_7) thus confirming the earlier findings (Maiti, Vairar, Singh et al. 1997) that NoOSP (October to next June) encouraged growth of natural weed flora (Table 1) including *Richardia* spp., *Euphorbia hitra*, and *Ageratum* spp. which harboured native VAM fungi and probably maintained already established mycelia networking extended from mycorrhizal roots as evidenced by Jasper, Abbott, and Robson (1989) and McGonigle, Evans, and Miller (1990) in maize. This led to a higher native VAM fungi population in NoOSP soils than OSP soils. In CS_3 , however, no added advantage of NoOSP was noticed because the long duration local variety (Laxmi) was harvested during February/March next year leaving only 3/2 months (March/April to June) for OSP. During those dry months, there could hardly be any effective tillage (ploughing) in OSP plots; therefore, detrimental effects of OSP on survival of VAM fungi remained void.

In CS_2 , toria as a sequence crop after rice (1994) reduced VAM fungi population because toria, a totally non-dependent crop on VAM, was not colonized by VAM fungi (Maiti, Variar, Singh, et al. 1997). Also, the subsequent (rice and horse gram) sequence was not suitable because of moisture stress; therefore, another cropping system (maize and niger) was tested. Shorter duration of maize (local variety) than rice (Vandana) allowed timely sowing of the sequence crop (niger) which could avoid acute moisture stress by getting established before the cessation of monsoon. Both the sequence cropping systems, however, led to substantial increase in native VAM fungi population with maximum increase in the latter sequence which also remained highest in June 1998 after the sole rice cropping in 1997. Farmer's practice of rotation (rice-fallow-black gram-rice), which was





Note $T_1: CS_1$ (rice sole [1993-1997]) + OSP (off-season ploughing); $T_2: CS_1 + NoOSP$, $T_3: CS_2$ (rice then toria [1993 and 1994], rice then horse gram [1995], maize then niger [1996], and rice sole [1997]) + OSP; $T_4: CS_2 + NoOSP$; $T_5: CS_3$ (rice and red gram intercropping [1993-97]) + OSP; $T_6: CS_3 + NoOSP$; $T_7: CS_4$ (rice sole [1994], fallow [1995], black gram [1996], and rice sole [1997]) + OSP; $T_8: CS_4 + NoOSP$

 Table 1
 Weed biomass during off-season (May) as influenced by agropractices

	Total weed	Total weed biomass during May (g/m²) ^f				
Treatments	1995	1996	1997			
T, ^g	98.8 ^d	106.2°	116.6 ^d			
T	382.6 ^b	456.6°	442.8 ^b			
$\mathbf{T}_{3}^{2_{i}}$	116.4 ^d	101.2°	109.2 ^d			
T [*] _j	443.8ª	523.6 ^b	577.6ª			
T [*] _E	227.0°	241.0 ^d	248.0°			
T ₁	231.0°	242.0 ^d	242.8°			
$T_{7}^{\circ}m$	113.7 ^d	247.2 ^d	114.2 ^d			
T ^{'n} ₈	381.9 ^b	674.1ª	439.9 ^b			

 f Mean of five samples; g CS₁ (rice sole [1993–97]) + OSP (off-season ploughing); h CS₁ + No OSP; i CS₂ (rice then toria [1993 and 1994], rice then horse gram [1995], maize then nigher [1996], and rice sole [1997]) + OSP; i CS₂ + NoOSP; k CS₃ (rice and red gram intercropping [1993–97]) + OSP; i CS₃ + NoOSP; m CS₄ (rice sole [1994], fallow [1995], black gram [1996], and rice sole [1997]) + OSP; r CS₄ + NoOSP

Note Means in each column followed by a common letter are not significantly different at the 5% level by DMRT

followed in the same order during 1994–97, increased VAM fungi population after black gram but reduced after rice sole, because of higher capability of black gram to harbour VAM fungi than rice (Maiti, Variar, and Saha 1995)

In all treatments, VAM fungi population peaked during September, not June (Figure 2) because of the presence of living root. Between rice and rice + red gram intercropping, the latter was found in earlier studies (Maiti, Variar, Singh, et al. 1997) to augment native VAM fungi population. When this was compared to maize sequence cropped by niger and rice sole in alternate years, the latter had a more encouraging effect on enhancement of native VAM fungi population in soil. Though the farmer's practice of crop rotation with NoOSP led to similar results, this practice should not be encouraged under the concept of intensive cultivation and achieving sustainable productivity from upland ecosystem.

The information about the effects of cropping system on native VAM fungi population would be useful when integrated with other crop management practices to realize higher productivity from rainfed uplands.

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Colonisation of upland rice by native VAM under rainfed monocropped ecosystem

In Recent Advances in Phytopathological Research, pp. 45–51, edited by A K Roy, and K K Sinha New Delhi: M D Publication

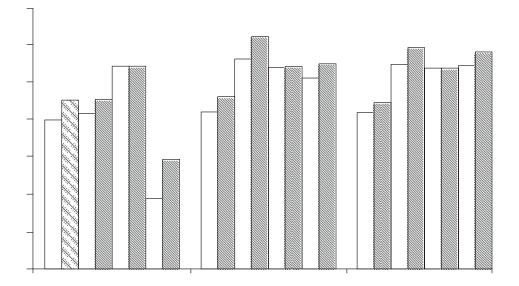


Figure 2 Effects of rice-based cropping systems on population dynamics of native VAM fungi at the end of crop season (September) as a function of time

Note $T_1: CS_1$ (rice sole [1993-1997]) + OSP (off-season ploughing); $T_2: CS_1 + NoOSP$, $T_3: CS_2$ (rice then toria [1993 and 1994], rice then horse gram [1995], maize then niger [1996], and rice sole [1997]) + OSP; $T_4: CS_2 + NoOSP$; $T_5: CS_3$ (rice and red gram intercropping [1993-97]) + OSP; $T_6: CS_3 + NoOSP$; $T_7: CS_4$ (rice sole [1994], fallow [1995], black gram [1996], and rice sole [1997]) + OSP; $T_8: CS_4 + NoOSP$

Maiti D, Variar M, Singh R K, Singh C V, Saha J. 1997 Enhancing VA-mycorrhizal association in rainfed upland rice (Oryza sativa L.) based cropping system Mycorrhiza News 8(4): 11–13

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New approaches

Estimation of total and living fungal biomass

Chitin and ergosterol contents of ectomycorrhizal roots were studied by ALF Ekblad, Wallander Hakan, and Nasholm Torgny (1998) in three sets of experiments to evaluate them as indicators of fungal biomass (The New Phytologist 138(1): 143-149). The first set of experiments showed that ageing has a marked effect on ergosterol concentration. The ergosterol content of sevenmonth-old, brown, shrunken, Pinus sylvestris -Paxillus involutus mycorrhizae was only 10 % of that found in white, turgid one- or four-month old specimens. This supports the hypothesis that the compound is a good indicator of living fungal biomass. Ageing has a lesser effect on chitin concentration since the chitin levels found in sevenmonth old mycorrhizae were 60% of the levels found in one- and four-month old specimens. Consequently, the chitin – ergosterol ratio increased from about 14-19 in one- and four-month-old mycorrhizae, respectively to about 110 in sevenmonth-old mycorrhizae.

In the second set of experiments, the authors found that variation in plant growth had no effect on the chitin–ergosterol ratio in whole root system of either *Alnus incana* or *P. sylvestris*, mycorrhizal with *Paxillus involutus*.

In the third set of experiments, the authors found a constant relationship between the two marker concentrations in 10-month old root system of *P. sylvestris* regardless of the fungal species involved, using *Paxillus involutus*, *Piloderma croceum* and *Suillus variegatus* as test organisms. Taken toPorter W M. 1979
The 'most probable number', method for enumerating infective propagules of vesiculararbuscular mycorrhizal fungi in soil
Australian Journal of Soil Research 17: 515-519

gether, the results of the study suggest that both chitin and ergosterol give reliable but different relative measures of fungal biomass in mycorrhizal roots. Furthermore, the authors demonstrated that in combination, the chemical markers can be used to estimate both total and living fungal biomass derived from the chitin–ergosterol ratio.

A simple staining technique for VAM fungi

A simple, reliable, and inexpensive staining technique was developed by H Vierheilig, A P Coughlan, U Wyss, Y Piche (1998) for staining vesicular-arbuscular mycorrhizal fungal colonization in root tissues (*Applied and Environmental Microbiology* 64(12): 5004–5007). Apart from application in research, this non-toxic, high-quality staining method could also be of great utility in teaching exercises.

Roots were cleared by boiling in 10% KOH. Boiling time differed according to the type of plant—bean, soybean, and maize for 5 minutes; cucumber, wheat, barley, and ryegrass roots for 3 minutes. The roots were then rinsed several times with tap water. Cleared roots were boiled for 3 minutes in a 5% ink-vinegar solution with pure white household vinegar (5% acetic acid). Roots were destained by rinsing in tap water (acidified with a few drops of vinegar) or by rinsing in pure vinegar. Destaining time differed according to the type of ink (for blue or black inks, destaining time was 20 minutes with acidified tap water, and for red ink, destaining time was 10 minutes with vinegar). The inks tested by the authors were purple, green, red, blue, and black.

Erratum

See Volume 12, Number 2, P. 21 under the column "New approaches" line 9–11: 'Proceedings of the second international conference on mycorrhiza, Uppsala, Sweden, 5–10 July, 1998' should read 'Proceedings of the first international conference on mycorrhizae, Berkeley campus, University of California, USA, 4–9 August, 1996'



List of cultures available with the Centre for Mycorrhizal Culture Collection

S No	CMCC Code	Culture Name	Host	s
1	EM-1001	Alpova diplophloeous	Pseudotsuga menziesii/	49
2	EM-1129	Alpova olivaceotinctus	Alnus rubra Abies concolor,	50 51
3	EM-1145	Amanita murina	Pinus jeffreyi Eucalyptus huidacsina	52 53 54
4	EM-1100	Amanita muscaria var. farmosa	bridgesina Populus tremuloides	55
5	EM-1084	Amanita muscaria	Pinus patula	56
6	EM-1069	Amanita muscaria	- <i>inus patata</i>	57
7	EM-1146	Amanita muscaria	Sous resineux	
8	EM-1060	Amanita muscaria	-	58
9	EM-1148	Boletus cavipes	-	59
10	EM-1277	Boletus cavipes	-	60
11	EM-1072	Boletus edulis	-	61
12	EM-1151	Cenococcum geophyllum	Picea	62
13	EM-1036	Cenococcum geophyllum	Tsuga	63
	TN <i>L</i> 4 6 6 7		mertensiana	
14	EM-1035	Cenococcum geophyllum	Pseudotsuga	64
15	EM-1150	Conservation and hullow	<i>mengiesii</i> Picea	65
15 16	EM-1150 EM-1149	Cenococcum geophyllum Cenococcum geophyllum	Tilleul	66
17	EM-1252	Cenococcum geophyllum Cenococcum geophyllum	Douglas	68
18	EM-1088	Cenococcum geophyllum	-	69
19	EM-1253	Chalciporus piperatus	-	70
20	EM-1153	Ectendomycorrhizae	Pinus sylvestris	71
21	EM-1154	Ectendomycorrhizae	Pinus sylvestris	72
22	EM-1155	Ectendomycorrhizae	-	73
23	EM-1157	Elaphomyces granulatus	Chene	74
24	EM-1156	Elaphomyces granulatus	Chene	75
25	EM-1108	Gautieria otthii	Pinus bankesiana	76
26	EM-1138	Gautieria caudata	Arbutus	
			menziesii,	78
			Pseudotsuga menziesii	79
27	AM-1005	Gigaspora margarita	-	
28	AM-1001	Glomus etunicatum	-	81
29	AM-1119	Glomus geosporum	-	
30	AM-1004	Glomus intraradices	-	82
31	AM-1006	Glomus mosseae	-	83
32	EM-1158	Gyrodon lividus	Alnus	84
33	EM-1163	Hebeloma crustuliniforme	Pin noir d' Autricha	0.7
34	EM-1160	crustulinijorme Hebeloma circinans	Picea	85
34 35	EM-1159	Hebeloma calyptosporum	Lisierie de pre	87
36	EM-1164	Hebeloma crustuliniforme	-	88
37	EM-1171	Hebeloma crustuliniforme	Douglus	89
38	EM-1172	Hebeloma cylindrosporum	-	90
39	EM-1173	Hebeloma cylindrosporum		91
40	EM-1170	Hebeloma crustuliniforme	Hetre	92
41	EM-1169	Hebeloma crustuliniforme		93
42	EM-1174	Hebeloma edurum	Picea	94
43	EM-1168	Hebeloma crustuliniforme		
44	EM-1165	Hebeloma crustuliniforme	-	95
$\frac{45}{46}$	EM-1175 EM-1008	Hebeloma edurum Hebeloma crustuliniforme		96
			monticola,	97
			Tsuga	98
47	EM 1100	T . L . L	heterophylla Diana	99
$\frac{47}{48}$	EM-1180 EM-1037	Hebeloma sinapizans Cenococcum geophyllum	Picea Tsuga	10
			mertensiana	10

S No	CMCC Code	Culture Name	Host
49	EM-1015	Hebeloma crustuliniforme	Pseudotsuga menziesii
50	EM-1183	Hebeloma truncatum	-
51	EM-1182	Hebeloma sinapizans	Hetre
52	EM-1184	Hebeloma vaccinum	-
53	EM-1176	Hebeloma ingratum	Tremble feuillus
54	EM-1177	Hebeloma mesophaeum	-
55	EM-1185	Hysterangium	Eucalyptus sp.
56	EM-1263	incarceratum Lactarius quietus	Chene
50 57	EM-1205 EM-1122	-	_
57	LIVI-1122	Laccaria farinacea	Tsuga mertensiana
50	EM 1102	I accouit hisslan	
58	EM-1103	Laccaria bicolor	Picea glauca
59	EM-1105	Laccaria laccata	Pinus resinosa
60	EM-1186	Laccaria amethystina	Sousdes
61	EM-1079	Laccaria laccata	Pinus patula
62	EM-1076	Laccaria laccata	Pinus patula
63	EM-1083	Laccaria fraterna	Eucalyptus
			globulus
64	EM-1086	Laccaria proxima	Betula
65	EM-1085	Laccaria laccata	Betula
66	EM-1187	Laccaria bicolor	-
67	EM-1188	Laccaria laccata	Picea sitchensis
68	EM-1193	Laccaria laccata	Sitka spruce
69	EM-1196	Laccaria tortilis	Bouleau
70	EM-1195	Laccaria proxima	-
71	EM-1102	Laccaria bicolor	Picea mariana
72	EM-1192	Laccaria laccata	Sitka spruce
73	EM-1189	Laccaria laccata	Ouescus ilex
74	EM-1190	Laccaria laccata	Hetre
75	EM-1191	Laccaria laccata	Eucalyptus sp.
76	EM-1104	Laccaria laccata	-
77	EM-1091	Laccaria amethystina	Pseudotsuga menziesii
78	EM-1058	Laccaria laccata	-
79	EM-1032	Laccaria laccata	Hemlock
80	EM-1032	Laccaria laccata	Pseudotsuga
80	EN1-1055	Laccana iaccaia	0
81	EM-1042	Laccaria laccata	menziesii Pseudotsuga
			menziesii
82	EM-1275	Laccaria proxima	Picea, Pinus
83	EM-1066	Laccaria laccata	-
84	EM-1031	Laccaria laccata	Pseudotsuga menziesii
85	EM-1067	Laccaria laccata	-
86	EM-1065	Laccaria laccata	-
87	EM-1090	Laccaria laccata	-
88	EM-1009	Laccaria laccata	Alpine
89	EM-1207	Laccinum scaber	Charme
90	EM-1207A	Laccinum scaber	Charme
91	EM-1261	Lactarius deliciosus	Pinus sylvestris
92	EM-1279	Lactarius hepaticus	-
93	EM-1264	Lactarius rufus	Sous resineux
94	EM-1052	Lactarius rufus	Larix larcina,
			Picea mariana
95	EM-1260	Lactarius chrysorrheus	Chene
96	EM-1200	Lactarius controversus	
			Populus euramericana
97	EM-1262	Lactarius deterrimus	-
98	EM-1206	Lactarius tabidus	-
99	EM-1199	Lactarius deterrimus	-
100	EM-1198	Lactarius controversus	Populus euramericana
101	EM-1259	Lactarius Chrysornheus	Hetre

S No	CMCC Code	Culture Name	Host	S No	CMCC Code	Culture Name
02	EM-1204	Lactarius subdulcis	Chene	154	EM-1044	Rhizopogon colossu
03	EM-1205	Lactarius subdulcis	Chene			1.8
04	EM-1202	Lactarius rufus	_			
05	EM-1203	Lactarius rufus	Sous resineux			
)6	EM-1133	Leccinum scabrum	Picea pungens,	155	EM-1038	Rhizopogon ellenae
	2	200000000000000000000000000000000000000	Betula	1.3.3	2002 10000	interpogen enema
)7	EM-1106	Leccinum insigne	Mixed			
'		Deceman insigne	hardwoodforest	156	EM-1134	Rhizopogon evader
3	EM-1281	Leccinum aurantiacum	Tremble	150	EM-1017	Rhizopogon occider
9	EM-1113		Pseudotsuga	158	EM-1017 EM-1049	Rhizopogon rubesce
	EM-1115	Martellia ellipsospora	menziesii	130	LINI-1049	Knizopogon rubesce
	EM 1120					
	EM-1130	Melanogaster	Arctostaphylos			
	T], <i>i</i> , <i>i</i>	tuberiformis	viscidia	1	FN <i>L</i> 4 A FA	
1	EM-1125	Melanogaster tuberiformi		159	EM-1053	Rhizopogon parksi
2	EM-1131	Melanogaster varigatus	Arctostaphylos	160	EM-1048	Rhizopogon
			viscidia			subcaerulescens
3	EM-1266	Paxillus involutus	-			
4	EM-1134	Paxillus involutus	Picea pungens,			
			Betula	161	EM-1054	Rhizopogon vulgar
5	EM-1209	Paxillus involutus	-			
6	EM-1270	Paxillus involutus	-	162	EM-1014	Rhizopogon vinicol
7	EM-1267	Paxillus involutus	Eucalyptus			10
			darlympleana			
8	EM-1268	Paxillus involutus	Chene	1		
9	EM-1269	Paxillus involutus	Tremble	163	EM-1050	Rhizopogon rubesco
9 0	EM-1073	Paxillus involutus	-	105	LIVI-1050	Knizopogon rubesci
			-			
1	EM-1208	Paxillus involutus	-			
2	EM-1141	Paxillus involutus	Castanea sativa	1.4	FN 1056	D1
3	EM-1217	Paxillus involutus	Chene	164	EM-1056	Rhizopogon villosu
4	EM-1212	Paxillus involutus	Populus			
			euramericana	165	EM-1028	Rhizopogon claviti
5	EM-1282	Paxillus involutus	Abies, Piecea			
6	EM-1055	Phaeolepiota aurea	Sitka spruce			
7	EM-1047	Phialocephala fortinii	Pinus	166	EM-1027	Rhizopogon hawke
8	EM-1219	Piloderma	-			
)	EM-1002	Pisolithus tinctorius	Tsuga			
			canadensis			
	EM-1057	Pisolithus tinctorius	_	167	EM-1007	Rhizopogon
	EM-1010	Pisolithus tinctorius	Loblolly pine,	1.0.	2001 1000	subcaerulescens
	200 1010	1 1301111113 111101011113	Pinus teada			var. subpannosus
	EM-1081	Pisolithus tinctorius	Eucalyptus	168	EM-1026	Rhizopogon
	LIVI-1001	1 isolilmus linelorius	tereticornis	100	LIVI-1020	ochraceorubens
3	EM-1059	Pisolithus tinctorius	lerencornis	169	EM-1018	Rhizopogon
			- Saw tooth Oak	109	LINI-1010	1 0
4	EM-1005	Pisolithus tinctorius	Saw tooth Oak Pin oak			subcaerulescens
5	EM-1006	Pisolithus tinctorius		170	EM 1110	
6	EM-1034	Pisolithus tinctorius	Loblolly pine	170	EM-1118	Rhizopogon vinicol
7	EM-1271	Pisolithus tinctorius	Pinus elliotii			
8	EM-1221	Pisolithus tinctorius	Chene			
9	EM-1223	Pisolithus tinctorius	Pinus caribaea	171	EM-1227	Rhizopogon luteolu
0	EM-1224	Pisolithus tinctorius	Chene	172	EM-1226	Rhizopogon luteolu
l	EM-1004	Pisolithus tinctorius	Virginia pines	173	EM-1128	Rhizopogon smithin
2	EM-1022	Rhizopogon subareolatus		174	EM-1127	Rhizopogon reaii
			menziesii	1		10
3	EM-1135	Rhizopogon elleane	Pinus ponderosa	1		
4	EM-1019	Rhizopogon fuscorubens	Pinus contorta	175	EM-1228	Rhizopogon nigresc
5	EM-1110	Rhizopogon vinicolor	-	176	EM-1220	Rhizopogon roseolu
, 5	EM-1025	Rhizopogon vulgaris	- Pinus	170	EM-1229	Rhizopogon vulgar
,	LIVI-102J	inizopogon ouigans	ponderosa,	177	EM-1232 EM-1231	Rhizopogon subrero
			-		EM-1231 EM-1230	
			Pseudotsuga	179		Rhizopogon rubesce
			menziesii,	180	EM-1012	Rhizopogon
_			Abies			ochraceorubens
7	EM-1043	Rhizopogon smithii	Pinus contorta	181	EM-1119	Rhizopogon occiden
8	EM-1040	Rhizopogon mutabilis	Pinus ponderosa	182	EM-1046	Rhizopogon ellenae
9	EM-1039	Rhizopogon vulgaris	Pinus ponderosa	1		
0	EM-1029	Rhizopogon cusickensis	Pinus	183	EM-1063	Rhizopogon vulgar
			ponderosa,	184	EM-1061	Rhizopogon rubesce
			Arbutus	185	EM-1011	Rhizopogon parksi
			menziesii			r-o Pariton
51	EM-1024	Rhizopogon	Tsuga,	1		
•	LIVI-1024	subcaerulescens	0,	1		
		subcuerniescens	Pseudotsuga	104	EM 1100	Dhizotan
	EM 1002	Diana di S	menziesii Di	186	EM-1109	Rhizopogon rubesce
	EM-1023	Rhizopogon vulgaris	Pinus contorta,	107	EV4 1114	D1' . '
	T1 <i>t t t t t t t t t t</i>		Abies amabilis	187	EM-1116	Rhizopogon vulgar
			Sitka spruce,	188	EM-1115	Rhizopogon
5	EM-1013	Rhizopogon arenicola	lodge pole pine			ochraceorubens

Culture Name	Host
Rhizopogon colossus	Pinus
Knizopogon colossus	ponderosa,
	Pseudotsuga
	menziesii
Rhizopogon ellenae	Arbutus
Knizopogon ellende	menziesii, Pinus
Phisopogon madama	ponderosa Pinus ponderosa
Rhizopogon evadens Rhizopogon occidentalis	Pinus ponderosa Pinus ponderosa
Rhizopogon rubescens	Pinus contorta,
Knizopogon rubescens	Abies
	lasiocarpa, Picea engelmani
Phisopogon banhaii	Picea engelmani Picea engelmani
Rhizopogon parksii Rhizopogon	Picea engelmant Pinus
subcaerulescens	
subcuerniescens	contorta, Abies
	lasiocarpa Picea engelmani
Rhizopogon vulgaris	-
Knizopogon vulgaris	Tsuga
Phineterson miniselen	mertensiana Taura
Rhizopogon vinicolor	Tsuga hatawat haulla
	heterophylla,
	Pseudotsuga
Phisobogon without	menziesii Dimus somtonta
Rhizopogon rubescens	Pinus contorta,
	Abies
	lasiocarpa,
DLinet and all and	Picea engelmani
Rhizopogon villosulus	Pseudotsuga
	menziesii
Rhizopogon clavitisporus	
	menziesii, Pinus
	ponderosa
Rhizopogon hawkerae	Pseudotsuga
	menziesii, Pinus
	ponderosa,
51	Abies
Rhizopogon	Abies grandis,
subcaerulescens	Pseudotsuga
var. subpannosus	menziesii
Rhizopogon	Pinus contorta
ochraceorubens	41. 1.
Rhizopogon	Abies grandis,
subcaerulescens	Pseudotsuga
D1	menziesii
Rhizopogon vinicolor	Mixed
	coniferous
B 1. 1 1	forest
Rhizopogon luteolus	-
Rhizopogon luteolus	-
Rhizopogon smithii	Shasta fir
Rhizopogon reaii	Pinus caribaea
	var.
D 11	hondurensis
Rhizopogon nigrescens	Pinus caribaea
Rhizopogon roseolus	Pinus
Rhizopogon vulgaris	-
Rhizopogon subrerolatus	Douglas
Rhizopogon rubescens	Pine
Rhizopogon	Pinus contorta,
ochraceorubens	Abies lasicarpa
Rhizopogon occidentalis	Pinus ponderosa
Rhizopogon ellenae	Pseudotsuga
	menziesii
Rhizopogon vulgaris	-
Rhizopogon rubescens	-
Rhizopogon parksii	Pseudotsuga
	menziesii, Pinus
	lambertianas,
	41. 1
	Abies concolor
Rhizopogon rubescens	Abies concolor Pinus
Rhizopogon rubescens	Pinus banksiana
Rhizopogon vulgaris	Pinus banksiana Pinus jeffreyi
Rhizopogon vulgaris	Pinus banksiana
	Pinus banksiana Pinus jeffreyi

Host

S No	CMCC Code	Culture Name	Host	S No	CMCC Code	Culture Name	Host
189	EM-1114	Rhizopogon hawkerae	Douglas fir	216	EM-1121	Suillus brevipes	Pinus contorta
190	EM-1112	Rhizopogon smithii	Pinus contorta	217	EM-1124	Suillus brevipes	Pinus contorta
191	EM-1126	Rhizopogon fuscorubens	Pinus radiata	218	EM-1087	suillus subluteous	Pinus Patula
192	EM-1140	Rhizopogon	Pinus ponderosa,	219	EM-1051	Suillus tomentosus	Pinus ponderosa
		ochraceisporus	Pseudotsuga	220	EM-1120	Suillus granulatus	Pinus contorta
		1	menziesii	221	EM-1021	Suillus tomentosus	Alder, Spruce,
193	EM-1142	Rhizopogon rubescens	Pinus nigra				lodgepole pine
194	EM-1143	Rhizopogon vinicolor	Pseudotsuga menziesii	222	EM-1117	Suillus punctatipes	Tsuga mertensiana,
195	EM-1132	Sarcodon scabrosus	Alnus sinnata				Abies lasiocarpa
196	EM-1273	Scleroderma aurantium	-	223	EM-1246	suillus variegatus	Sous resineux
197	EM-1235	Scleroderma flavidum	Eucalyptus	224	EM-1245	suillus variegatus	Sous resineux
		2	camaldulencis	225	EM-1247	Thelephora terrestris	-
198	EM-1233	Scleroderma cepa	Eucalyptus sp.	226	EM-1077	Thelephora terrestris	Pinus patula
199	EM-1107	Scleroderma citrinum	Hardwood	227	EM-1062	Thelephora terrestris	-
			forest	228	EM-1249	Tricholoma populinum	Populus
200	EM-1241	Suillus grannulatus	-				euramericana
201	EM-1240	Suillus granulatus	Pinus sylvestris	229	EM-1248	Tricholoma albobruneum	Pine
202	EM-1030	Suillus americanus	Pinus strobus	230	EM-1250	Tricholoma scalpturatum	Populus
203	EM-1075	Suillus variegatus	-				euramericana
204	EM-1243	Suillus luteus	-	231	EM-1041	Tuber melanosporum	-
205	EM-1244	Suillus luteus	Pinus sylvestris	232	EM-1161	Hebeloma crustuliniformi	Picea
206	EM-1237	suillus bellini	-	233	EM-1078	Thelephora terrestris	Betula
207	EM-1238	suillus bovinoides	Pinus sylvestris	234	EM-1276	Tricholoma ustale	Hetre
208	EM-1236	suillus bellini	-	235	EM-1144	Wilcoxinia mikolae	-
209	EM-1074	Suillus bovinus	-	236	EM-1283	Scleroderma verucosum	Sal
210	EM-1239	suillus bovinus	Sous resineux	237	EM-1284	Scleroderma cepa	P. wallichiana
211	EM-1136	suillus luteus	Pinus	238	EM-1285	Lycoperdon Sp.	P. wallichiana
212	EM-1137	suillus luteus	Picea	239	EM-1286	Russula Sp.	P. smithiana
213	EM-1071	Suillus luteus	-	240	EM-1287	Suillus Sp.	Pinus caribaea
214	EM-1111	Suillus lakei	Douglas fir, Hemlock	241	EM-1288	Geastrum Sp.	Eucalyptus tereticornis
215	EM-1123	Suillus tomentosus	Pinus monticola, Tsuga heterophylla	242	EM-1289	Pisolithus tinctorius	-

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Conferences, congresses, seminars, symposiums, and workshops

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Edewageningen, The Netherlands 20 November 1 December 2000	International Symposium. Durable resistance: key to sustainable agriculture Dr J E Parlevliet, Laboratory of Plant Breeding, Wageningen University, PO Box 386, NL 6700, Wageningen, The Netherlands
	Fax +31 317 483457 • E-mail jan.par,evo,et@users.pv.wau.nl Web http://www.spq.wau.nl/pv/symposium.htm
Belém , Brazil 4-8 December 2000	Integrated Management of Neotropical Rain Forests by Industries and Com- munities. Dr Natalino Silva, Brazilian Agricultural Research Corp, CP 48, CEP 66240 Belem, Para, Brazil
	Fax 55 91 2269845 • Tel 55 91 2266622 • E-mail natalino@cpatu.embrapa.br

Canberra , Australia 10–13 December 2000	5th Pacific Rim Bio-based Composites Symposium Philip Evans, Department of Forestry, Australian National University, Canberra ACT 0200 Australia
	<i>Fax</i> 61 2 6249 0746 • <i>Tel</i> 61 2 6249 3628 • <i>E-mail</i> Bio.symposium@anu.edu.au <i>Web</i> http://online.anu.edu.au/Forestry/wood/bio/bio.html
Fremantle, Australia 18-25 April 2001	16th Commonwealth Forestry Conference Libby Jones, UK Forestry Commission, 231 Corstorphine Road, Edinburgh EH 12 7AT, UK
	Fax 44 (0) 131 334 0442 • Tel 44 (0) 131 314 6137 • E-mail libby.jones@forestry.gov.uk
Austria June 2001	 FAO/ECE/ILO Workshop on New Developments of Wood Harvesting with Cable Systems. R. Heinrich, Forest Harvesting, Trade and Marketing Branch, Forest Products Division, FAO, Viale delle Terme di Caracalla, 00100 Rome, Italy
	Fax 39 06 5705 5137 • E-mail Forest-Harvesting@FAO.org
Yogyakarta, Indonesia 11-13 June 2001	International Conference on ex situ and in situ Conservation of Commercial Tropical Trees. Ms Soetitah S. Soedojo, ITTO Project PD 16/96 Rev.4 (F), Faculty of Forestry,
	Gadjah Mada University, Bulaksumur, Yogyakarta 55281, Indonesia.
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	Fax 1 970 498 1027 • Tel 1 970 498 1012 • E-mail mryan@lamar.colostate.edu
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	Fax 1 541 737 1393 • Tel 1 541 737 6558 • E-mail strauss@fsl.orst.edu Web http://www.cof.orst.edu/cof/extended/conferen/treebio/
Aberdeen , Scotland 12-14 September 2001	Dynamics of Forest Insect Populations Dr Andrew Liebhold, USDA Forest Service, Northeastern Forest Experiment Station, Forestry Sciences Laboratory, 180 Canfield St., Morgantown West Virginia 26505, USA
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Sydney , Australia 9–14 September 2001	5th International Flora Malesiana Symposium. Contact: Dr Barry Conn, Royal Botanic Gardens Sydney, Mrs Macquaries Road, Syd- ney NSW 2000, Australia,
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Valdivia, Chile October 2001	Improvement and Culture of Eucalypts. IUFRO 2.08.03. Contact: Dr. Roberto Ipinza, Universidad Austral de Chile, PO Box 1241, Valdivia, Chile
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